# Research article

# Defining Neotropical otter *Lontra longicaudis* distribution, conservation priorities and ecological frontiers

# Marcelo Lopes Rheingantz<sup>1</sup>, Jorge Fernando Saraiva de Menezes<sup>2</sup> and Benoit de Thoisy<sup>3</sup>

<sup>1</sup>Laboratório de Ecologia e Conservação de Populações - Av. Carlos Chagas Filho, 373 - Centro de Ciências da Saúde, Instituto de Biologia. Departamento de Ecologia, Sala A027, Cx. P. 68020, Universidade Federal do Rio de Janeiro. Rio de Janeiro, RJ, CEP 21941-590. Brazil, <sup>2</sup>Laboratório de Vida Selvagem – Embrapa - Rua 21 de setembro, 1880. Bairro N.S. Fátima CxP 109. Corumbá, MS, CEP 79320-900. Brazil, <sup>3</sup>Association Kwata, BP 672, 97335 Cayenne cedex, French Guiana Corresponding author and Email address: Marcelo Lopes Rheingantz mlrheingantz@gmail.com

#### Abstract

Understanding a species' occurrence requirements is essential for its conservation, and species distribution models (SDMs) are a powerful tool for this purpose. Here we estimated a SDM based on actual distribution information, in relation to climatic, hydrological, human population, and vegetation data sets, to understand the ecological requirements and geographic distribution of the Neotropical otter *Lontra longicaudis*, a species whose habitat requirements and conservation needs are mostly unknown. Using MaxEnt, we defined its potential distribution and most suitable areas to indicate priority areas for research and to analyze the efficiency of Protected Areas (PAs). Our findings suggest that the range of Neotropical otters could extend beyond their present estimated distribution, adding new areas in northeastern Brazil, Andean region, west Ecuador, Venezuela, Peru, Mexico, and Argentina, with higher suitability in rain forests (especially Atlantic and Amazon Forests). We also found that PAs are the most suitable areas for otter distribution. Although better than non-protected areas, PAs are close to the median of the suitability values, indicating that they still can be improved to conserve otters. Annual temperature and human population density explained most data variance in our model. We suggest the change of the actual status of Neotropical otter to Least Concern or Near Threatened categories. We recommend verifying the possible sympatry with other otters, and demonstrate that rudimentary and/or occasional recent data of occurrence can also be used in SDMs and contribute to species conservation.

Keywords: distribution range, Neotropical otter, niche modeling, reserve design, suitability

#### Resumo

Compreender os requisitos para a ocorrência de uma espécie é fundamental para sua conservação, e os modelos de distribuição de espécies (SDM) são uma ferramenta poderosa para essa finalidade. Aqui nós estimamos um SDM a partir de informações reais de distribuição, em relação a conjuntos de dados climáticos, hidrológicos, de população humana e de vegetação para entender as exigências ecológicas e a distribuição geográfica da lontra Neotropical Lontra longicaudis, espécie cujos requisitos de habitat e necessidades para a sua conservação são praticamente desconhecidos. Usando MaxEnt, definimos sua distribuição potencial e as áreas mais adequadas para a espécie a fim de indicar áreas prioritárias para pesquisa e analisar a eficácia das áreas protegidas (APs). Nossos resultados indicam que a distribuição da lontra neotropical poderia se estender além de sua distribuição atual estimada, adicionando novas áreas no nordeste do Brasil, na região andina, no oeste do Equador, na Venezuela, no Peru, no México e na Argentina, com maior adequabilidade em florestas tropicais (especialmente Mata Atlântica e Floresta Amazônica). Descobrimos também que APs são as áreas mais adequadas para a distribuição de lontra. Porém, embora sejam mais adequadas do que áreas não protegidas, as APs estão perto da mediana dos valores de adequabilidade, o que indica que elas ainda podem ser melhoradas para melhor conservar lontras. Temperatura anual e Densidade populacional humana foram as variáveis que melhor explicaram a variância dos dados em nosso modelo. Sugerimos a mudança do status real de lontra neotropical para Pouco Preocupante ou categorias Quase Ameaçada. Recomendamos verificar a eventual simpatria com outras lontras, e demonstramos que mesmo dados rudimentares e/ou ocasionais de ocorrência também podem ser usado em SDMs e contribuir para a preservação de espécies.

Palavras-chave: adequabilidade, desenho de reservas, distribuição geográfica, lontra Neotropical, modelagem de nicho

Received: 21 Febraury 2014; Accepted 28 April 2014; Published: 23 June 2014

**Copyright:** © Marcelo Lopes Rheingantz, Jorge Fernando Saraiva de Menezes and Benoit de Thoisy. This is an open access paper. We use the Creative Commons Attribution 3.0 license http://creativecommons.org/licenses/by/3.0/us/. The license permits any user to download, print out, extract, archive, and distribute the article, so long as appropriate credit is given to the authors and source of the work. The license ensures that the published article will be as widely available as possible and that your article can be included in any scientific archive. Open Access authors retain the copyrights of their papers. Open access is a property of individual works, not necessarily journals or publishers.

Cite this paper as: Marcelo Lopes Rheingantz, Jorge Fernando Saraiva de Menezes and Benoit de Thoisy. 2014. Defining Neotropical otter *Lontra longicaudis* distribution, conservation priorities, and ecological frontiers. *Tropical Conservation Science* Vol.7 (2): 214-229. Available online: <a href="https://www.tropicalconservationscience.org">www.tropicalconservationscience.org</a>

## Introduction

Understanding drivers of species distribution is essential to its conservation and determining ecological requirements[1,2]. Although field techniques can address several questions about species conservation as well as or even better than estimated distributions, experiment costs are prohibitive for large mammals. The Neotropical otter (*Lontra longicaudis*), a solitary and elusive semiaquatic carnivore [3] with one of the largest home ranges of Lutrinae [4], is a species whose large scale dynamics cannot be addressed by traditional methods. Despite all efforts, present data are not accurate enough to determine the conservation status of the Neotropical otter based on International Union for Conservation of Nature (IUCN) criteria, and the species is now labeled as 'data deficient' [5]. In order to correct such knowledge deficiencies, the IUCN Otter Specialist Group (during the XIth IUCN OSG International Otter Colloquium in Pavia, Italy in 2011) recommended that research on this species, currently one of the least studied of the subfamily Lutrinae [5,6], should focus on defining its actual distribution, population status, and habitat requirements.

The status of Neotropical otter is important to the conservation of river ecosystems because of its role as a top-chain predator on those ecosystems, and its presence can help maintain local biodiversity [3,7]. In addition, the species requires large territory extensions to establish a viable population, as each individual requires dozens of kilometres of riparian habitats [8–10] and also depends on water physicochemical conditions and habitat structures to persist [8,9]. Modifications of those characteristics can considerably affect otter populations [9], reducing local diversity.

Despite its importance to species conservation, species distribution areas are frequently estimated using a minimum convex polygon, a method that ignores species' ecological constraints [1,2]. A more recent set of techniques, known as species distribution models (SDMs), can overcome this limitation. SDMs also provide extra information in several areas of ecology, evolution, conservation, and management [11,12]. SDMs were initially developed to predict distributions from incomplete data [13], but have been widely used in other areas related to conservation, such as the impact of biological invasions (see[14]), threatened species monitoring (see [15]), estimating distribution expansion of recovering species [16,17], and evaluating climate change impacts [17–19]. These models also identify the most suitable areas for a species, from which we can infer where it is most abundant [20,21] and which environmental characteristics are important to it.

In species with large distribution areas, such as the Neotropical otter, that knowledge can be further used to prioritize regions for species conservation while considering economic interest [22]. Large species also have more accurate representations of their distribution, as most models are based on climatic variables, which are more determinant at larger scales [23,24].

Among SDMs, Maximum entropy (MaxEnt) is a method used for predictions or inferences with presence-only data [21,25–28], and is as efficient as models with both presence and absence data [26,29]. This model tool is highly recommended for studies with the same goals as our work, since the model discriminates between appropriate and inappropriate areas fairly well when compared with other methods [26,30]. Considering all these advantages, coupled with the lack of information regarding Neotropical otter, we developed a questionnaire to collect information about *Lontra longicaudis* occurrences and sent it to researchers and environmentalists. Using their information, we modeled the species distribution using MaxEnt.

In summary, we intended to: (a) obtain an updated database with all available Neotropical otter locations in all its distribution range; (b) estimate the species distribution; (c) identify climatic, environmental and population variables that most influenced *Lontra longicaudis* habitat suitability; (d) identify recommended areas for future studies; and (e) verify if the actual network of Protected Areas is more suitable for Neotropical otter than non-protected areas.

# **Methods**

To test the hypotheses above, we adopted the following method/work flow: first a database of known records and abiotic variables was assembled; then, we used MaxEnt Modeling to create a species distribution model; and last, specific procedures were used to address each hypothesis.

#### Database description

We surveyed current Neotropical otter occurrences to identify where the species is present, sampling literature and asking researchers for otter records to estimate species potential distribution.

We included all geo-referenced records published in scientific papers (search engines: Google Scholar, ISI Web of Knowledge, Scopus, Scielo) and provided by specialists working on faunal inventories, especially semiaquatic mammals (consulting 231 members of the IUCN Otter Specialist Group), and sent 331 requests directly to known scientists who work with river or mammal ecology within *Lontra longicaudis* distribution. We considered both direct (visual, captures, camera-traps), and indirect observations (including footprints, spraints and/or hair). We considered as presences only data with date information, exact geographic coordinates (converted in WGS-84 system), and collector's name. All occurrence data used in this work are listed in the supplementary material. We obtained data from Brazil, Argentina, Uruguay, Paraguay, Peru, Ecuador, Colombia, French Guiana, Suriname, Venezuela, Mexico, Bolivia, Costa Rica, and Panama, for a period covering 1991 to 2012. With this time span of 30 years we assumed that our record database accurately describes the current distribution of the Neotropical otter (Fig. 1).

We obtained abiotic data from different sources. To build the SDM, we chose *a priori* variables that we assumed to be the most significant to the species and/or were used in previous Neotropical mammal studies [28,31]: annual precipitation, precipitation of the driest month, precipitation seasonality, precipitation of the warmest quarter, annual temperature, isothermallity, and temperature standard. Other variables deemed less important were not included, as including more variables reduces model fitness [32]. One data source was the

Worldclim database, formed by 18 global raster maps representing climatic variables, averaged between years 1950-2000 [33]. We included altitude from the Shuttle Radar Topography Mission (SRTM) digital elevation raster map, human population density in 2000, obtained from GPWv3 [34], and vegetation cover in 2012, measured as the NDVI from MODIS sensors [35]. We also derived a raster map of percentage of water bodies from SRTM Water Body vector Data [36]. The value of a cell equals how many of the 100 sub-cells were covered by water body vectors. All maps had their resolution downgraded to 0.2 decimal, the minimum resolution available to all. All data, independent of source, were clipped between northern Mexico to northern Argentina, the frontiers of the species' current IUCN distribution (Fig. 1).

To test the suitability of existing Protected Areas (PAs) (see below), we used the vector map WDPA Conservation Units [37]. This global database is a joint venture by UNEP-WCMC and IUCN World Commission on Protected Areas in collaboration with governments, NGOs, academia and industry. These data are constantly updated.

# Model estimation / assessment of the quality of the model

Once the database was complete, we generated a Kernel density map (0.01555 bandwidth), considering 95% of distribution use. The Kernel was used to ensure that background sampling only included areas where otter could exist (a SDM requirement) [21,25]. Within this interval, we selected 1,000 random locations to be used as background data for modeling.

MaxEnt 3.3.3k [25] estimates the probability distribution of the maximum entropy of each environmental variable within the study area. This distribution is calculated with the constraint that the expected value of each environmental variable under it matches the empirical average generated from environmental values associated with species occurrence data [25]. When MaxEnt is applied to presence-only SDM, the pixels of the study area make up the space on which the MaxEnt probability distribution is defined; pixels with known species occurrence records constitute the sample points; and the features become environmental variables. We used 75% of the records for training the model and 25% for test. The best parameter estimate was calculated using 5,000 interactions with bootstrap replication. We set the convergence threshold of  $1.0 \times 10^{-5}$  (i.e. increase in model fitness below which the model stops). In addition, the model was replicated 15 times to calculate confidence intervals, and model fitness was assessed based in the area under the curve [38].

#### Hypothesis-specific procedures

After identification of the most suitable areas for the occurrence of otters, a series of transformations were required to address each hypothesis.

To describe otter potential distribution, we first generated a consensus map from the 15 MaxEnt models using the weighted average method [39], which uses the AUC value of each model as its weight. We then considered two approaches to derive binary presence—absence maps from the continuous consensus map of suitability: one more restrictive and conservative, maximizing the sum of sensitivity and specificity (max SSS) (more details in [40]); and a second, less restrictive, with the lowest predicted value threshold (LPV) (see [25,28,41]). LPV is the lowest value of environmental suitability within an occurrence record, and Max SSS is an objective method that optimizes the discrimination between presence/absence in the same way as that between presence/random points.

Second, to test the efficiency of PAs, we compared the distribution of suitability values within and outside PAs and ran a Kolmogorov-Smirnov (KS) test to verify if these two curves differ significantly.

Last, to recommend regions for future studies and conservation initiatives, we reasoned that places with high suitability for otter and far from previous studies have a better chance to be different ecosystems/biomes and to have different environmental characteristics. Therefore, a study in those areas can explain if the otter uses different habitats in ways similar to those shown in previous studies. To accomplish this, we multiplied the suitability of a pixel by its distance to the nearest record, based on the assumption that ecological similarities decrease with distance. The index was then divided by its highest value, to ensure it varies across the more intuitive scale of 0 to 1. Note that this index implies that distance and suitability are equally important, and therefore it may recommend areas with low suitability.

All analyses were performed, and figures made, in ArcMap 9.3 [42]. The Kolmogorov-Smirnov test was conducted in the R Software [43].

#### **Results**

We obtained information on Neotropical otter in 14 of the countries that encompassed its historical distribution, including 565 occurrences (available at: http://goo.gl/G6BaqG). Our study included new information about Neotropical otter distribution, with new occurrences recorded outside the recognized distribution ranges (Fig. 1 and [5,44,45]). Those new occurrences included areas in northeastern Brazil [46], Venezuela, Mexico, and also in the Pacific portions of Ecuador, Peru and Colombia.

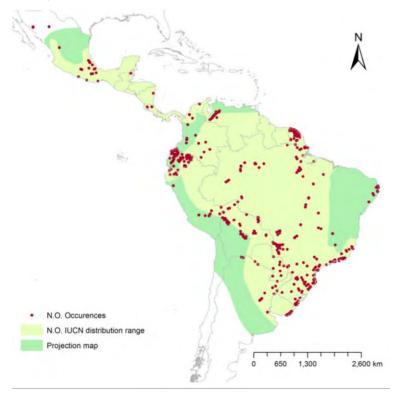


Fig. 1. Map showing Neotropical otter (*Lontra longicaudis*) unique records (n=565), in red, overlaid with IUCN distribution, in light green, and Maxent projection, in green

#### Neotropical otter potential distribution

We generated a map with *L. longicaudis* potential distribution, based on the two threshold criteria. With the LPV criterion we obtained a broader map than with the Max SSS criterion (Fig. 2). AUC average value was 0.809±0.016. Neotropical otter distribution range, according to the LPV, totalled 17.056.351 km², which represents 45.4% of North, Central and South Americas' total area and is 34% greater than IUCN's historical otter distribution. Considering Max SSS, we obtained 8.148.746 km², which is 21.7% of the area of the three Americas and 63.6% of IUCN's historical distribution. The consensus map, according to the LPV, contained 25.8% (4.410.023 km²) of the distribution outside the IUCN historical range, while the Max SSS map contained 9.8% (799.728 km²). 1.2% (163.834 km²) of the IUCN distribution map was outside the LPV consensus map, and 42.6% (5.460.145 km²) was outside of the Max SSS consensus map (Fig. 2).

The only regions within IUCN distribution, but absent according to LPV, were areas outside kernel regions in Central America. Distribution according to Max SSS was smaller than IUCN historical range in almost all countries (Fig. 2).

Ranges generated by both thresholds and inside the minimum convex polygon containing all locations were much wider than those indicated in B1 criteria for any IUCN endangered status (more details on [47]), where maximum value of extent of occurrence to be listed as vulnerable is less than 20,000 km<sup>2</sup>.

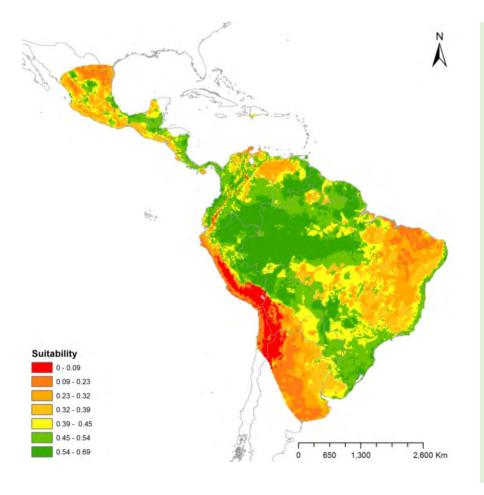
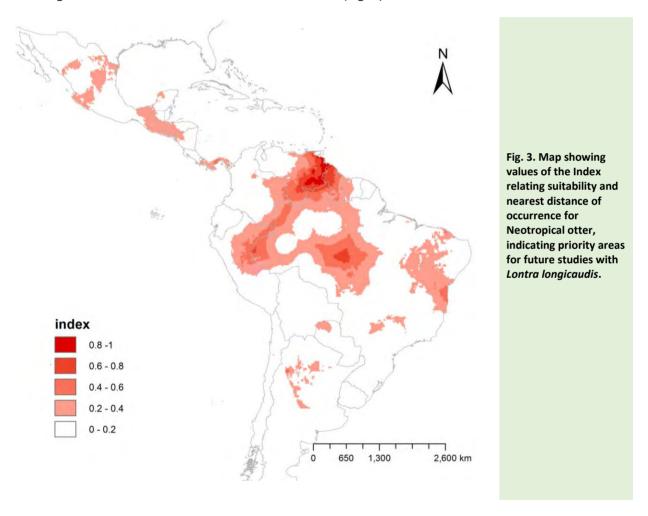


Fig. 2. Map showing potential geographical distribution and suitability values for Neotropical otter (Lontra longicaudis) accordingly with MAXENT. Suitability values above the lowest predicted value threshold (LPV) are shown from orange to green. Values above the maximum sum of sensitivity and specificity threshold (Max SSS) are green and values below the LPV threshold are red.

Regarding the importance of each variable, annual temperature (60.5%) was the one that most contributed to the model: highest values of suitability were associated with higher annual temperatures. Human population density was identified as the second contribution variable (29.9%), and was inversely related to otter suitability. However, higher human population densities were not completely unsuitable for otters (suitability=0.55 in low densities and 0.42 in higher population densities). Other variables had less than 2% contribution each. Vegetation, proportion of water, and altitude contributed 0%, 0.2% e 0.2% respectively.

#### Priority areas for future studies with Lontra longicaudis

The index proposed areas with higher values and recommended for future studies, to be the frontier between Venezuela and Guyana, other Venezuelan areas, northeastern Brazil, and regions within the Brazilian and Peruvian Amazon (Fig. 3).



# Suitability of Protected Areas vs Non-Protected Areas

Areas inside PAs (number of cells=7432) showed higher suitability values than areas outside them (number of cells=29,976) (KS=0.3775 p<0.0001). Kernel densities of PA were higher in higher suitabilities (0.4-0.65, with peak around 0.58) than non-protected areas (0.1-0.5, with peak around 0.4) (Fig. 4).

Beyond PAs, the higher values were observed in the Amazon region, which was the region with higher suitability cells. Other areas that had PAs with good suitability were the coastal region of the Atlantic Forest, the Guyana shield, the Pantanal, and many regions of Central America.

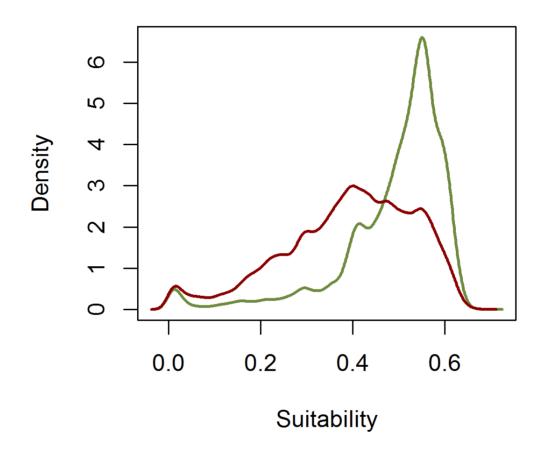


Fig. 4. Distribution of *Lontra longicaudis'* suitability estimated by kernel in protected (green line) and unprotected (red line) areas. The densities are related with the number of pixels with this value of suitability.

#### **Discussion**

In this study, we tentatively propose a model for Neotropical otter *Lontra longicaudis* (Figure 5) distribution. SDMs are powerful statistical techniques that can be used to describe species' geographic ranges [48]. There are several methods to model species distribution, but the most commonly used now is the maximum entropy approach (Maxent; [25]). The high average AUC indicates that our models had good predictive power and can be used to analyze environmental suitability. Our results showed that different variables were important to the potential distribution of this top-chain semiaquatic mammal, as shown before with another otter[49] and with another large Neotropical carnivore, the jaguar[28]. Annual temperature and human population density were the most important predictors in our model for Neotropical otter in a large scale. Our model was innovative in two aspects: we demonstrated that PAs have more suitable areas for *L. longicaudis* than non-protected areas, and we identified the Peruvian and Brazilian Amazon, the borders between Venezuela and Guyana, and Northeastern Brazil as areas for future studies.

Our model indicated that areas with high annual temperatures are the most suitable, similar to what was described in the species' revision[44]. The other important variable was human population density. Although suitability decreased with human density, its value was high even in human-dense areas. These results appear to be consistent with the biology of various otter species, as otters can occur in degraded and/or urban areas [3,50–53]. However, NDVI small contribution to the model did not support another previous hypothesis: that *L. longicaudis* requires good riparian vegetation and good den availability [54]. This difference between previous studies and our work is probably due to scaling, as we used a larger, less accurate scale (0.2 decimal degree cell size) to identify the fine scale of microhabitat selection. In microhabitat scale, there are evidences of Neotropical otter using more forested areas to select their holts [55,56]. Another environmental variable previously suggested as important, but not supported by our model, is altitude. We had locations from sea level to 4,200m high (e.g. Peruvian Andes), with many occurrences above 1,500m or below 300m from sea level. Despite that range, altitude did not influence most of our model replicates, in disagreement with previous studies that described Neotropical otter less abundant at high altitudes [44].

Considering the above variables, the Amazon, Atlantic Forest, and Pantanal are the most suitable environments for *L. longicaudis*, a challenge for Neotropical otter conservation. The Atlantic Forest is one of the most threatened biomes of the world[57,58], and the Amazon, although well preserved in comparison with others, has been severely degraded in recent years [59,60], with almost 20% of its Brazilian land cover (60% of total) converted to land use[61]. Despite being less suitable, regions in Central America may justify the importance of investments in them as there are well-preserved areas in several countries, and previous studies have indicated the continent as a priority for carnivore conservation [62].

Categorizing the suitability values to estimate Neotropical otter distribution also provided novel insights. Distributions categorized by both LPV and Max SSS overlap substantially with distributions described before (see [5,44]). However, even the distribution according to Max SSS, the most restrictive one, included areas absent in IUCN's map. This shows that the IUCN map needs to be updated. New regions to be added include northeastern Brazil, the Andes (in Peru, Ecuador and Colombia), northern Mexico, western Colombia, and Bolivia. These new areas may provide valuable information about the species as well as insight on what really limits the species' occurrence. The range increase beyond the IUCN northern limit seems to be related to an increase in resolution. Models based in minimum convex polygon are coarser than our distribution modeling (for comparisons with our map, see [5,44,63]). This expansion

also indicates that *L. longicaudis* may expand its distribution to southern United States and thus overlap with *L. canadensis* distribution [64,65]. Further studies to identify this potential interaction or the barrier that separates both species are recommended. Another contribution of categorizing the suitability values, is to assess in which IUCN category the species should be allocated. The estimated species distribution is beyond 20,000 km², in spite of the threshold criteria used. Consequently, the species cannot be considered vulnerable or any other category below, according to the IUCN species extent, and we suggest the species should be moved to either Near Threatened or Least Concern.

In addition to estimating Neotropical otter distribution, we also compared the environmental suitability inside and outside PAs. In our study we showed that PAs are significantly more suitable and can help in Neotropical otter conservation, even though they have not been created specifically to protect otters. The higher suitability values inside the PAs are not surprising, as many of them were designed based on river basins or river courses [66,67].

PAs, mainly larger ones, reduce human influence and can retain populations and also assemblages [68]. PAs are especially important to conserve Amazonian ecosystems, as the majority of the remaining forests throughout the Amazon basin are within protected areas and only 1.5% of these PAs have been deforested [69]. Tropical PAs are ecologically linked to their surrounding habitats, so it's important to conserve not only the area within them, but also surrounding areas in order to maintain local biodiversity [70].



Fig. 5. Neotropical otter *Lontra longicaudis* in river margins of Brazilian Pantanal. Photo credits: Caroline Leuchtenberger/Instituto Federal Farroupilha, with permission.

# Implications for conservation

Our analysis shows that spatial distribution models (SDMs) [71] can be used to evaluate whether the PAs are really conserving *L. longicaudis* within them. Such approaches increase the utility of SDMs, which were already used in PA design and management[20,72,73], to guide population surveys [74–76] even when no information about species' absences was available. Regarding future studies, we recommend focusing on northeastern Brazil, as it represents one of driest areas in Neotropical otter distribution. The Amazon regions in Brazil,

Guyana, and Venezuela and the Pantanal are also identified as informative regions due to their low human density, mostly warm climate, and well-preserved areas.

As lack of knowledge about a species' distribution is one of the most important issues in mammal conservation [77], we suggest gathering and compiling otter occurrence data in Argentina, Suriname, Ecuador, Central America, Mexico, and Northeastern Brazil to corroborate our model predictions that these regions are suitable for the species. Despite Neotropical otters' wide distribution range and their being well documented in most of their distribution, our study indicates that new research is necessary to obtain Neotropical otter occurrences in areas with little information.

Despite some undocumented areas, our results suggest that we can move *L. longicaudis* from Data Deficient to Least Concern or Near Threatened Categories. We demonstrate here that rudimentary and/or occasional data of occurrence can also be used in spatial distribution models and contribute to species conservation by better describing their distribution.

# **Acknowledgements**

We thank the IUCN OSG, specially the chair of the group Nicole Duplaix, for allowing the contact with researchers as an official task and for providing information. We also thank to Jordi Ruiz-Olmo and Luiz Gustavo Oliveira-Santos for suggestions in the manuscript and Bernardo Araújo for the final review. We had the support of Graduate Program in Ecology of the Universidade Federal do Rio de Janeiro and PIBIC/UFRJ. This work would not have been possible without all of the Survey respondents who took the time to share their knowledge for the benefit of the species: Andrés Pautasso, Laura Fasola, Claudio Chehébar and Marcelo Cassini (Argentina); Robert Wallace (WCS/Bolívia), Arturo Muñoz and Aidan Maccormick (Noel Kempff Mercado Museum/Bolivia); Roberta Elize Silva, Patrícia Farias and Fernando Rosas (INPA/Brazil); Miriam Marmontel, Joana Silva Macedo, Henrique Lazzarotto and Danielle Lima (Instituto Mamirauá/ Brazil); Carlos Henrique Salvador, Everton Bernardo de Miranda, Pamela Antunes, Carlos André Zucco, Diogo Loretto and Luiz Gustavo Oliveira-Santos (UFRJ/Brazil); Caroline Leuchtenberger and Carolina Ribas (Embrapa Pantanal/Brazil); Helen Waldemarin and Vera de Ferran (Ecology and Environment do Brasil/Brazil); Manoel Comes Muanis (Associação Ecológica Ecomarapendi/Brazil); Tiago Gomes dos Santos (Universidade Federal do Pampa/Brazil); Fernanda R. Rizzoto and Jaime Martinez (Universidade de Passo Fundo/Brazil); Vinicius Galvão Bastazini, Fernando Margues Quintela, Luciane Dutra Coletti and Carlos Benhur Kasper (UFRGS/Brazil); Jorge José Cherem, Mauricio Graipel, Henrique Krauser, Felipe Fantacini and Barbara Carpegianni (UFSC/Brazil); Marcelo Arasaki (ONG MAE/Brazil); Cláudia Cristina de Sousa de Melo (Museu Paraense Emilio Goeldi, Brazil); Silvana Campello (Instituto Araguaia/Brazil), Oldemar Carvalho-Junior (Insituto Ekko Brasil/Brazil); Daniel Louzada-Silva (Secretaria de Educação DF/Brazil); Beatriz Beisiegel and Livia Rodrigues (CENAP ICMBio/Brazil); Marcelo Labruna (USP/Brazil); Julio Cesar Voltolini (Univ. Taubaté/Brazil); José Lailson Brito Junior (UERJ/Brazil); Marcelo F. G. Brito (Univ. Federal Sergipe/Brazil); Marcelo Passamani (Univ. Federal Lavras/Brazil); David Costa Braga (Brazil); Maria Piedad Baptiste, Fernando Trujillo, Lida Marcela Franco Perez, Maria Fernanda Cely Garcia and Juan Carlos Botello (Colombia); Victor Utreras and Mario Quevedo (WCS Ecuador), Felix Manging (Universidad Guayaquil/Ecuador); Juan Pablo Gallo-Reynoso (IUCN OSG/Mexico); José L. Cartes, Hugo Del Castillo and Alberto Yanosky (Asociación Guyra/Paraguay); Rob Williams (Sociedad Zoologica de Francfort/Peru); and Ildemaro González (IUCN OSG/Venezuela).

#### References

- [1] Brown, J. H. 1995. *Macroecology*. 270. University of Chicago Press, Chicago.
- [2] Brown, J. H., Stevens, G. C. and Kaufman, D. M. 1996. The geographic range: size, shape, boundaries, and internal structure. *Annual Review of Ecology and Systematics* **27:** 597–623.
- [3] Kruuk, H. 2006. *Otters: Ecology, Behaviour and Conservation*. 336. Oxford University Press, Oxford.
- [4] Rosas, F. C. W. 2004. Lontra, *Lontra longicaudis* (carnivora: mustelidae). In: *História Natural, Ecologia e Conservação de Algumas Espécies de Plantas e Animais da Amazônia* 330. Instituto de Pesquisa da Amazonia, Manaus.
- [5] Waldemarin, H. F. and Alvarez, R. 2008. *Lontra longicaudis*. http://www.iucnredlist.org/details/12304/0
- [6] Rheingantz, M. L., Waldemarin, H. F., Rodrigues, L. and Moulton, T. P. 2011. Seasonal and spatial differences in feeding habits of the neotropical otter *Lontra longicaudis* (carnivora: mustelidae) in a coastal catchment of southeastern brazil. *Zoologia* (*Curitiba*, *Impresso*) **28:** 37–44.
- [7] Estes, J. a, Terborgh, J., Brashares, J. S., Power, M. E., Berger, J. et al. 2011. Trophic downgrading of planet Earth. *Science* **333**: 301–6.
- [8] Green, J., Green, R. and Jefferies, D. 1984. A radio-tracking survey of otters *Lutra lutra* on a perthshire river system. *Lutra* 27: 85–147.
- [9] Bowyer, R. T., Testa, J. W. and Faro, J. B. 1995. Habitat selection and home ranges of river otters in a marine environment: effects of the Exxon Valdez oil spill. *Journal of Mammalogy* **76:** 1–11.
- [10] Blundell, G. M., Maier, J. A. K. and Debevec, E. M. 2014. Linear home ranges: effects of smoothing, sample size, and autocorrelation on kernel estimates. *Ecological Monographs* 71: 469–489.
- [11] Guisan, A. and Thuiller, W. 2005. Predicting species distribution: offering more than simple habitat models. *Ecology Letters* **8**: 993–1009.
- [12] Zimmermann, N. E., Edwards, T. C., Graham, C. H., Pearman, P. B. and Svenning, J. C. 2010. New trends in species distribution modelling. *Ecography* **33**: 985–989.
- [13] Fisher, D. and Dickman, C. 1993. Body size-prey relationships in insectivorous marsupials: tests of three hypotheses. *Ecology* **74:** 1871–1883.
- [14] Oliveira, M. D., Hamilton, S. K., Calheiros, D. F., Jacobi, C. M. and Latini, R. O. 2000. Modeling the potential distribution of the invasive golden mussel *Limnoperna fortunei* in the upper paraguay river system using limnological variables. *Brazilian Journal of Biology* **70:** 831–840.
- [15] Osborne, P. E., Alonso, J. C. and Bryant, R. G. 2001. Modelling landscape-scale habitat use using GIS and remote sensing: a case study with great bustards. *Journal of Applied Ecology* **38**: 458–471.
- [16] Corsi, F., Duprè, E. and Boitani, L. 1999. A large-scale model of wolf distribution in Italy for conservation planning. *Conservation Biology* **13**: 150–159.
- [17] Cianfrani, C., Le Lay, G., Hirzel, A. H. and Loy, A. 2010. Do habitat suitability models reliably predict the recovery areas of threatened species? *Journal of Applied Ecology* **47:** 421–430.
- [18] Colombo, a. F. and Joly, C. A. 2010. Brazilian atlantic forest *lato sensu*: the most ancient brazilian forest, and a biodiversity hotspot, is highly threatened by climate change. *Brazilian journal of biology = Revista brasleira de biologia* **70**: 697–708.
- [19] Cassini, M. H. 2011. Ranking threats using species distribution models in the IUCN red list assessment process. *Biodiversity and Conservation* **20**: 3689–3692.

- [20] Traill, L. W. and Bigalke, R. C. 2006. A presence-only habitat suitability model for large grazing african ungulates and its utility for wildlife management. *African Journal of Ecology* **45**: 347–354.
- [21] Clement, L., Catzeflis, F., Richard-Hansen, C., Barrioz, S., de Thoisy. 2014. Conservation interest of spatial distribution modelling applied to large vagile mammals. *Tropical Conservation Science* 7(2) 192-213.
- [22] Rondinini, C. and Boitani, L. 2007. Systematic conservation planning and the cost of tackling conservation conflicts with large carnivores in Italy. *Conservation biology* **21**: 1455–62.
- [23] Pearson, R. G., Dawson, T. P. and Liu, C. 2004. Modelling species distributions in britain: a hierarchical integration of climate and land-cover data. *Ecography* **27**: 285–298.
- [24] Luoto, M., Virkkala, R. and Heikkinen, R. K. 2007. The role of land cover in bioclimatic models depends on spatial resolution. *Global Ecology and Biogeography* **16:** 34–42.
- [25] Phillips, S., Anderson, R. and Schapire, R. 2006. Maximum entropy modeling of species geographic distributions. *Ecological Modelling* **190**: 231–259.
- [26] Wisz, M. S., Hijmans, R. J., Li, J., Peterson, A. T., Graham, C. H. et al. 2008. Effects of sample size on the performance of species distribution models. *Diversity and Distributions* **14:** 763–773.
- [27] Riordan, E. C. and Rundel, P. W. 2009. Modelling the distribution of a threatened habitat: the California sage scrub. *Journal of Biogeography* **36:** 2176–2188.
- [28] Tôrres, N. M., De Marco, P., Santos, T., Silveira, L., de Almeida Jácomo, A. T. et al. 2012. Can species distribution modelling provide estimates of population densities? a case study with jaguars in the neotropics. *Diversity and Distributions* **18**: 615–627.
- [29] Hernández, L., Laundré, J. W. and Gurung, M. 2005. From the field: use of camera traps to measure predation risk in a puma mule deer system. *Wildlife Society Bulletin* **33**: 353–358.
- [30] Elith, J., Graham, C. H., Anderson, R. P., Dudík, M., Ferrier, S. et al. 2006. Novel methods improve prediction of species' distributions from occurrence data. *Ecography* **2**: 129–151.
- [31] DeMatteo, K. E. and Loiselle, B. a. 2008. New data on the status and distribution of the bush dog (*Speothos venaticus*): evaluating its quality of protection and directing research efforts. *Biological Conservation* **141**: 2494–2505.
- [32] Elith, J. and Leathwick, J. R. 2009. Species distribution models: ecological explanation and prediction across space and time. *Annual Review of Ecology, Evolution, and Systematics* **40**: 677–697.
- [33] Hijmans, R. J., Cameron, S. E., Parra, J. L., Jones, P. G. and Jarvis, A. 2005. Very high resolution interpolated climate surfaces for global land areas. *International Journal of Climatology* **25**: 1965–1978.
- [34] CIESIN and CIAT 2005. Gridded population of the world, version 3 (gpwv3): population density grid. <a href="http://sedac.ciesin.columbia.edu/data/set/gpw-v3-population-density">http://sedac.ciesin.columbia.edu/data/set/gpw-v3-population-density</a>
- [35] NASA Land Processes Distributed Active Archive Center (LP DAAC). 2012. Terra/modis net primary production yearly I4 global 1km. https://lpdaac.usgs.gov/products/modis products table/mod17a3
- [36] Farr, T., Rosen, P. A., Caro, E., Crippen, R., Duren, R. et al. 2007. The shuttle radar topography mission. *Review of Geophysics* **45**: 1–43.
- [37] IUCN and UNEP. 2009. The world database on protected areas (wdpa). http://www.wdpa.org/
- [38] Bradley, A. 1997. The use of the area under the ROC curve in the evaluation of machine learning algorithms. *Pattern recognition* **30:** 1145–1159.

- [39] Marmion, M., Parviainen, M., Luoto, M., Heikkinen, R. K. and Thuiller, W. 2009. Evaluation of consensus methods in predictive species distribution modelling. *Diversity and Distributions* **15:** 59–69.
- [40] Liu, C., White, M. and Newell, G. 2013. Selecting thresholds for the prediction of species occurrence with presence-only data. *Journal of Biogeography* **40:** 778–789.
- [41] Liu, C., Berry, P. M., Dawson, T. P. and Pearson, R. G. 2005. Selecting thresholds of occurrence in the prediction of species distributions. *Ecography* **28**: 385–393.
- [42] ESRI. 2008. Arcgis 10. Environmental Research Institute, Redlands, CA.
- [43] R Core Team. 2013. R: a language and environment for statistical computing. R Core Team, Vienna, Austria.
- [44] Larivière, S. 1999. Lontra longicaudis. Mammalian Species **609**: 1–5.
- [45] Emmons, L. H. and Feer, F. 1997. *Neotropical Rainforest Mammals: a Field Guide*. 396. University of Chicago Press, Chicago.
- [46] Astúa, D., Asfora, P. H., Aléssio, F. M. and Langguth, A. 2010. On the occurrence of the neotropical otter ( *Lontra longicaudis* ) ( Mammalia: Mustelidae ) in northeastern Brazil. *Mammalia* 74: 213–217.
- [47] IUCN. 2010. *IUCN Red List of Threatened Species*. version 2010.4. http://www.iucnredlist.org
- [48] Hortal, J., Lobo, J. and Jiménez-Valverde, A. 2012. Basic questions in biogeography and the (lack of) simplicity of species distributions: putting species distribution models in the right place. *Natureza & Conservação* **10**: 108–118.
- [49] Cianfrani, C., Lay, G. Le, Maiorano, L., Satizábal, H. F., Loy, A. et al. 2011. Adapting global conservation strategies to climate change at the european scale: the otter as a flagship species. *Biological Conservation* **144**: 2068–2080.
- [50] Pardini, R. and Trajano, E. 2008. Use of shelters by the neotropical river otter (lontra longicaudis) in an Atlantic Forest stream, southeastern Brazil. *Journal of Mammalogy* **80:** 600–610.
- [51] Alarcon, G. G. and Simões-Lopes, P. C. 2003. Preserved versus degraded coastal environments: a case study of the neotropical otter in the environmental protection area of Anhatomirim, southern Brazil. *IUCN Otter Specialist Group Bulletin* **20**: 1–10.
- [52] Rheingantz, M. L., Oliveira-santos, L. G., Waldemarin, H. F. and Caramaschi, E. P. 2012. Are otters generalists or do they prefer larger, slower prey? feeding flexibility of the neotropical otter *Lontra longicaudis* in the Atlantic Forest. *IUCN otter specialist bulleting* **29:** 80–94.
- [53] Macdonald, S. M. and Mason, C. F. 1980. Observations on the marking behaviour of a coastal population of otters. *Acta Theriologica* **25**: 245–253.
- [54] Blacher, C. 1987. Ocorrência e preservação de lutra longicaudis (Mammalia mustelidae) no literal de Santa Catarina. Boletim FBCN 22: 105–117.
- [55] Quadros, J. and Monteiro-Filho, E. 2002. Sprainting sites of the neotropical otter, *Lontra longicaudis*, in an Atlantic Forest area of southern Brazil. *Mastozoología neotropical* **9:** 39–46.
- [56] Kasper, C. and Bastazini, V. 2008. Trophic ecology and the use of shelters and latrines by the neotropical otter (*Lontra longicaudis*) in the Taquari valley, southern Brazil. *Iheringia. Série ...* **98:** 469–474.
- [57] Ranta, P., Blom, T., Niemelä, J., Joensuu, E. and Siitonen, M. 1998. The fragmented Atlantic rain forest of Brazil: size, shape and distribution of forest fragments. *Biodiversity and Conservation* **7:** 385–403.

- [58] Ribeiro, M. C., Metzger, J. P., Martensen, A. C., Ponzoni, F. J. and Hirota, M. M. 2009. The Brazilian Atlantic Forest: how much is left, and how is the remaining forest distributed? Implications for conservation. *Biological Conservation* **142:** 1141–1153.
- [59] Shukla, J., Nobre, C. and Sellers, P. 1990. Amazon deforestation and climate change. *Science* **247**: 1–4.
- [60] Rosa, I. M. D., Purves, D., Souza, C. and Ewers, R. M. 2013. Predictive modelling of contagious deforestation in the brazilian Amazon. *PloS one* **8:** e77231.
- [61] Pereira, D., Santos, D., Vedoveto, M., Guimarães, J. and Veríssimo, A. 2010. *Fatos Florestais Da Amazônia 2010*. 124. Instituto do Homem e Meio Ambiente da Amazônia, Belém.
- [62] Loyola, R. D., Oliveira-Santos, L. G. R., Almeida-Neto, M., Nogueira, D. M., Kubota, U. et al. 2009. Integrating economic costs and biological traits into global conservation priorities for carnivores. *PloS one* **4:** e6807.
- [63] Eisenberg, J. and Redford, K. 1999. *Mammals Of The Neotropics. V.3. The Central Neotropics: Ecuador, Peru, Bolivia, Brazil.* The University of Chicago Press, Chicago.
- [64] Polechla, P. 1990. Action plan for north american otters. In: Otters: An action plan for their conservation Foster-Turley, P., Macdonald, S. and Mason, C.(Eds.) 126. International Union for the Conservation of Nature, Gland, Switzerland.
- [65] Larivière, S. 1998. Lontra canadensis. Mammalian Species **587**: 1–8.
- [66] Balana, B. B., Yatich, T. and Mäkelä, M. 2011. A conjoint analysis of landholder preferences for reward-based land-management contracts in kapingazi watershed, eastern mount Kenya. *Journal of environmental management* **92**: 2634–46.
- [67] Shanahan, S. A., Nelson, S. M., Van Dooremolen, D. M. and Eckberg, J. R. 2011. Restoring habitat for riparian birds in the lower Colorado river watershed: an example from the las vegas wash, nevada. *Journal of Arid Environments* **75:** 1182–1190.
- [68] Cantú-Salazar, L. and Gaston, K. J. 2013. Species richness and representation in protected areas of the western hemisphere: discrepancies between checklists and range maps. *Diversity and Distributions* **19**: 782–793.
- [69] Numata, I. and Cochrane, M. A. 2012. Forest fragmentation and its potential implications in the brazilian Amazon between 2001 and 2010. *Open Journal of Forestry* 2: 265–271.
- [70] Laurance, W. F., Useche, D. C., Rendeiro, J., Kalka, M., Bradshaw, C. J. a et al. 2012. Averting biodiversity collapse in tropical forest protected areas. *Nature* 489: 290–4.
- [71] Guisan, A. and Zimmermann, N. E. 2000. Predictive habitat distribution models in ecology. *Ecological Modelling* **135**: 147–186.
- [72] Loiselle, B. A., Howell, C. A., Graham, C. H., Goerck, J. M., Brooks, T. et al. 2003. Avoiding pitfalls of using species distribution models in conservation planning. *Conservation Biology* **17**: 1591–1600.
- [73] Wilson, K. A., Westphal, M. I., Possingham, H. P. and Elith, J. 2005. Sensitivity of conservation planning to different approaches to using predicted species distribution data. *Biological Conservation* 22: 99–112.
- [74] Bourg, N. A., McShea, W. J. and Gill, D. E. 2005. Putting a cart before the search: successful habitat prediction for a rare forest herb. *Ecology* **86**: 2793–2804.
- [75] De Siqueira, M. F., Durigan, G., de Marco Júnior, P. and Peterson, A. T. 2009. Something from nothing: using landscape similarity and ecological niche modeling to find rare plant species. *Journal for Nature Conservation* **17:** 25–32.
- [76] Jackson, C. R. and Robertson, M. P. 2011. Predicting the potential distribution of an endangered cryptic subterranean mammal from few occurrence records. *Journal for Nature Conservation* **19:** 87–94.

[77] Rondinini, C., Rodrigues, A. S. L. and Boitani, L. 2011. The key elements of a comprehensive global mammal conservation strategy. Philosophical transactions of the Royal Society of London. Series B, Biological sciences **366**: 2591–7.